

TECHNICAL NOTE

Intrinsic Sorption Potential of Aluminum-Substituted Calcium Silicate Hydroxy Hydrate for Cesium-137

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The sorption capacity of synthetic calcium silicate hydrate $Ca_5Al_xSi_{6-x}O_{18}H_{24}\cdot nH_2O$ (where x = 0.025) for Cs^{137} has been investigated at 25°C using the radiometric method. Cs137 uptake at different concentrations has been investigated under uniform conditions at various time intervals. The material exhibits the capacity to uptake Cs137 from tracer solutions. Cs sorption has been studied with pure Cs solution and at 1000 to 5000 times concentration of sodium ions as the competing cation. Advanced Cement Based MATERIALS 1995, 2, 80-83

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everal cement based formulations have been suggested for immobilization of nuclear waste [1-5]. The exact mechanism of immobilization is, however, complex and still unknown. Long-term management of nuclear waste involves conversion of Cs and Sr into aluminosilicate in suitable matrix [6–8]. The interpretation of cation uptake in cement matrix becomes rather difficult because the uptake is not exclusively due to the clay or zeolite type purely thermodynamic cation exchange process. It is also associated with the simultaneous precipitation of some complex silicate species due to the highly alkaline nature of hydrated cement mixtures. In the recent past, several articles have appeared with thermodynamic and kinetic data on uptake reaction of various cations onto the cement phase which is identical to mineral 11-Å tobermorite, a calcium silicate hydrate close to Ca₅Si₆O₁₈H₂·nH₂O. Komarneni and Roy [3] have reported its Cs selectivity in the presence of various competing alkali metal and alkaline earth cations. Shrivastava and Glasser [9] have investigated its uptake behavior for several divalent cations. Use of naturally occurring clay minerals like vermiculite has been extensively studied with actual rad wastes. Vermiculite has shown remarkably higher pick-up for radioactive Cs [10]. The reports on Cs selectivity of 11-Å tobermorite have prompted us to evaluate this material for its potential in immobilization of Cs137. The radiochemical data on Cs uptake may also be useful in solving the complex mechanism of ion exchange and cation uptake reactions of tobermorite in particular and other calcium silicate minerals in general.

Experimental

Phase-pure tobermorite appropriate composition (98 to 99% pure) has been synthesized using the hydrothermal method described by Kalousek [11]. The calculated amounts of lime, silica, and alumina were weighed to keep the molar ratio Ca:(Si + Al) equal to 0.8. The products have been characterized by chemical analysis, powder X-ray diffraction between glancing angle $2\theta =$ 5 to 80° using Ni-filtered CuK radiation on a Phillips PW-1820 diffractometer. The morphology of tobermorite crystals has been examined with a Phillips 420 TEM microscope. The Cs uptake at different concentrations was followed by the radiometric method: the inactive CsNO₃ solution was spiked with Cs¹³⁷ tracer, a β and γ emitter with a half-life of 30 years. The pickup kinetics were studied after contacting the exchangers with the above solutions for different time periods, and liquid aliquots were withdrawn, centrifuged, and counted for Cs137. The clear solution free from partic-

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ulate matter was counted using a Tl-activated NaI scintillation counter.

Results and Discussion

The phase-pure Al-substituted 11-Å tobermorite was characterized by its typical powder X-ray diffraction pattern, which matches in intensities and positions with the JCPDS file [12]. The resulting product gives 22 reflections with D values between 113 and 11.5 nm. The product appeared to be free of unreacted Ca(OH)2, SiO₂, and CaCO₃. On shaking the sample in water (solid:water = 1:100), the solution gives a pH of 8.5. The electron micrograph shows a high content of platy crystals characteristic of tobermorites [13]. Crawford et al. [14] have reported that the ordinary Portland cement (OPC) has negligible sorption potential for Cs. The hydrated cements have solid mass with complex microporous regimes. The poor sorption potential of OPC for Cs is attributed to the absence of any crystallochemical incorporation of Cs in cement gels. Studies were made to investigate the pick-up behavior of 11-Å tobermorite for Cs¹³⁷ from pure and mixed cationic solutions at room temperature. The study records the sorption potential of 11-Å tobermorite for Cs¹³⁷ in pure and mixed cationic solutions (Na⁺ + Cs⁺), respectively. The presence of competing Na+ ions in higher concentration lowers the Cs uptake capacity and the

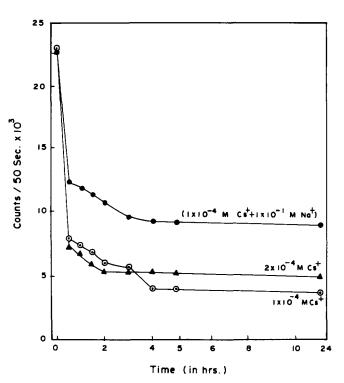


FIGURE 1. Depletion of γ activity of Cs¹³⁷ with time.

selectivity coefficients to a sizable extent. Tobermorite exhibited 26% Cs pick-up in the presence of 5000 times concentrated Na+ solution after 24 hours of equilibration, whereas the pickup in pure CsNO3 solution is as

TABLE 1. Radiochemical data on concentration dependence of Cs¹³⁷ uptake onto 2.5 mol% Al-substituted tobermorite

| Equilibration time (hours) | Molarity of Solution (M) | | | | | | | | | | |
|----------------------------|--|--------------------|-------------------|--|-----------------|-----------------|----------------|------|--|--|--|
| | | 2×10^{-4} | M Cs ⁺ | 1.5 × 10 ⁻⁴ M Cs ⁺ | | | | | | | |
| | Initial count rate (counts per 50 seconds) | | | | | | | | | | |
| | | 22,905 : | ± 50 | 20,400 ± 87 | | | | | | | |
| | Δ | % Cs removal | K _d | Df | Δ | % CS removal | K _d | Df | | | |
| 0.5 | 15,714 | 69 | 218.52 | 3.19 | 8485 | 41 | 355.14 | 1.71 | | | |
| 1.0 | 16,203 | 71 | 241.76 | 3.42 | 10,972 | 52 | 581.88 | 2.16 | | | |
| 1.5 | 16,981 | 74 | 286.65 | 3.87 | 10,749 | 52 | 556.88 | 2.11 | | | |
| 2.0 | 17,481 | 76 | 322.29 | 4.22 | 11,734 | 5 <i>7</i> | 677.01 | 2.35 | | | |
| 3.0 | 17,481 | 76 | 322.29 | 4.22 | 11, 7 98 | <i>57</i> | 685.77 | 2.37 | | | |
| 4.0 | 16,481 | 76 | 322.29 | 4.22 | 12,188 | 59 | 742.08 | 2.48 | | | |
| 5.0 | 17,481 | 76 | 322.29 | 4.22 | 13,712 | 66 | 1025.12 | 3.05 | | | |
| 24.0 | 17,481 | 76 | 322.29 | 4.22 | 13,372 | 65 | 981.34 | 2.90 | | | |
| Volume of solution (ml) | | 100 | | | | 500 | | | | | |
| Amount of exchanger (g) | | | | | | | | | | | |

 $[\]Delta$ = Initial counts - final counts (per 50 seconds).

[%] Cs removal = (No. of counts sorbed onto exchanger)/(Initial counts in solution) \times 100.

 K_d = Selectivity coefficient (mg⁻¹) = (Co/Ce₋₁) A/M, where Co = initial No. of counts; Ce = equilibrium counts in solution; A = volume of solution; and M mass of exchanger.

Df = Decontamination factor = (initial counts)/(final counts).

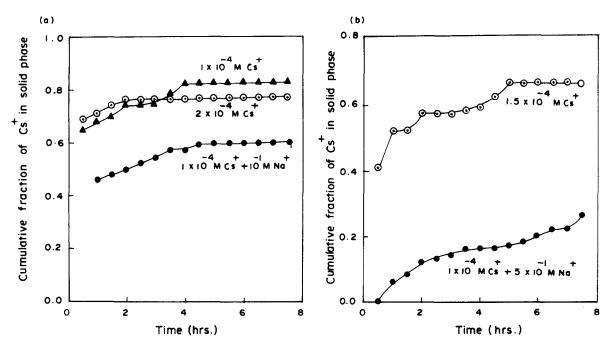


FIGURE 2. Cumulative fraction of Cs exchanged on 2.5 mol% Al-substituted tobermorite.

high as 82%. Table 1 records the time and concentration dependence of Cs uptake in different solutions. Figure 1 compares the depletion of γ activity due to sorption of radiocesium onto tobermorite with and without competing cations. Figure 2 compares the cu-

mulativefraction of Cs in pure solutions with that of mixed cationic solutions containing Na⁺ and Cs⁺ together. The sorption properties may be attributed to the crystallochemical incorporation of Cs in a layered lattice framework of tobermorite [15–20]. The Cs up-

TABLE 1. Continued

| Molarity of Solution (M) | | | | | | | | | | | | | |
|--|---|--|--|---|--|---|--|--|--|--|--|--|--|
| 1.0 × 10 ⁻⁴ M Cs ⁺ | | | $1.0 \times 10^{-4} \text{ M Cs}^+ + 0.1 \text{ M Na}^+$ | | | | $1.0 \times 10^{-4} \text{ M Cs}^+ + 0.5 \text{ M Na}^+$ | | | | | | |
| Initial count rate (counts per 50 seconds) | | | | | | | | | | | | | |
| 22,905 ± 50 | | | 22,905 ± 50 | | | | 25,414 ± 75 | | | | | | |
| % Cs removal | K _d | Df | Δ | % Cs removal | K _d | Df | Δ | % Cs removal | K _d | Df | | | |
| 65 | 189.27 | 2.89 | 10,501 | 46 | 84.66 | 1.85 | 1566 | 0.6 | 32.88 | 1.07 | | | |
| 68 | 209.36 | 3.09 | 10,984 | 48 | 92.14 | 1.92 | 1988 | 08 | 42.43 | 1.08 | | | |
| 70 | 233.50 | 3.34 | 11,489 | 50 | 100.64 | 2.01 | 3050 | 12 | 68.19 | 1.14 | | | |
| 74 | 278.47 | 3.78 | 12,013 | 52 | 110.29 | 2.10 | 3150 | 13 | 70.74 | 1.14 | | | |
| 7 5 | 297.72 | 3.97 | 13,148 | 57 | 134.75 | 2.35 | 3833 | 16 | 88.80 | 1.18 | | | |
| 82 | 460.16 | 5.60 | 13,411 | 59 | 141.26 | 2.41 | 4019 | 16 | 93.92 | 1.19 | | | |
| 82 | 460.16 | 5.60 | 13,411 | 59 | 141.26 | 2.41 | 4482 | 18 | 107.06 | 1.21 | | | |
| 82 | 460.16 | 5.60 | 13,411 | 59 | 141.26 | 2.41 | 6567 | 26 | 174.22 | 1.35 | | | |
| 100 | | | | 100 | | | | 500 | | | | | |
| | 22,905 : % Cs removal 65 68 70 74 75 82 82 82 | 22,905 ± 50 % Cs removal K _d 65 189.27 68 209.36 70 233.50 74 278.47 75 297.72 82 460.16 82 460.16 82 460.16 | $ \begin{array}{c ccccccccccccccccccccccccccccccccccc$ | 1.0 × 10^{-4} M Cs ⁺ 1.0 × Initial council cou | 1.0 × 10^{-4} M Cs ⁺ Initial count rate (countrate) 22,905 ± 50 22,905 ± % Cs removal % Cs removal 65 189.27 2.89 10,501 46 68 209.36 3.09 10,984 48 70 233.50 3.34 11,489 50 74 278.47 3.78 12,013 52 75 297.72 3.97 13,148 57 82 460.16 5.60 13,411 59 82 460.16 5.60 13,411 59 82 460.16 5.60 13,411 59 82 460.16 5.60 13,411 59 | 1.0 × 10^{-4} M Cs ⁺ + 0.1 M N Initial count rate (counts per 50 s 22,905 ± 50 22,905 ± 50 % Cs removal K _d Df Δ % Cs removal K _d 65 189.27 2.89 10,501 46 84.66 68 209.36 3.09 10,984 48 92.14 70 233.50 3.34 11,489 50 100.64 74 278.47 3.78 12,013 52 110.29 75 297.72 3.97 13,148 57 134.75 82 460.16 5.60 13,411 59 141.26 82 460.16 5.60 13,411 59 141.26 82 460.16 5.60 13,411 59 141.26 82 460.16 5.60 13,411 59 141.26 82 460.16 5.60 13,411 59 141.26 | 1.0 × 10^{-4} M Cs ⁺ + 0.1 M Na ⁺ Initial count rate (counts per 50 seconds) 22,905 ± 50 22,905 ± 50 % Cs removal K _d Df Δ % Cs removal K _d Df 65 189.27 2.89 10,501 46 84.66 1.85 68 209.36 3.09 10,984 48 92.14 1.92 70 233.50 3.34 11,489 50 100.64 2.01 74 278.47 3.78 12,013 52 110.29 2.10 75 297.72 3.97 13,148 57 134.75 2.35 82 460.16 5.60 13,411 59 141.26 2.41 82 460.16 5.60 13,411 59 141.26 2.41 82 460.16 5.60 13,411 59 141.26 2.41 82 460.16 5.60 13,411 59 141.26 2.41 82 460.16 5.60 13,411 59 141.26 2.41 <td>1.0 × 10^{-4} M Cs⁺ + 0.1 M Na⁺ 1.0 : Initial count rate (counts per 50 seconds) 22,905 ± 50 % Cs removal K_d Df Δ % Cs removal K_d Df Δ 65 189.27 2.89 10,501 46 84.66 1.85 1566 68 209.36 3.09 10,984 48 92.14 1.92 1988 70 233.50 3.34 11,489 50 100.64 2.01 3050 74 278.47 3.78 12,013 52 110.29 2.10 3150 75 297.72 3.97 13,148 57 134.75 2.35 3833 82 460.16 5.60 13,411 59 141.26 2.41 4019 82 460.16 5.60 13,411 59 141.26 2.41 4482 82 460.16 5.60 13,411 59 141.26 2.41 4482</td> <td>$\begin{array}{c ccccccccccccccccccccccccccccccccccc$</td> <td>$\begin{array}{c ccccccccccccccccccccccccccccccccccc$</td> | 1.0 × 10^{-4} M Cs ⁺ + 0.1 M Na ⁺ 1.0 : Initial count rate (counts per 50 seconds) 22,905 ± 50 % Cs removal K _d Df Δ % Cs removal K _d Df Δ 65 189.27 2.89 10,501 46 84.66 1.85 1566 68 209.36 3.09 10,984 48 92.14 1.92 1988 70 233.50 3.34 11,489 50 100.64 2.01 3050 74 278.47 3.78 12,013 52 110.29 2.10 3150 75 297.72 3.97 13,148 57 134.75 2.35 3833 82 460.16 5.60 13,411 59 141.26 2.41 4019 82 460.16 5.60 13,411 59 141.26 2.41 4482 82 460.16 5.60 13,411 59 141.26 2.41 4482 | $ \begin{array}{c ccccccccccccccccccccccccccccccccccc$ | $ \begin{array}{c ccccccccccccccccccccccccccccccccccc$ | | | |

take of tobermorite is also important from the viewpoint of structurally related materials which crystallize as a result of the hydrothermal reaction of lime and silica. The examination of published thermodynamic and kinetic data on sorption of other alkali metal, alkaline earth, and transition metal cations reveals that the uptake in tobermorite is partly due to the ion exchange, along with the simultaneous precipitation of complex silicate species in alkaline medium and the breaking of bonds from the edge of the layered lattice [21]. The characterization of the end products of the sorption or exchange reactions is in progress. The exchange is said to have been caused by the Al3+ substitution for Si⁴⁺, that is, H₃O⁺ subsequently participating in the exchange with Cs ions [10,15]. The simultaneous proportionate release of Ca²⁺ ions in solution to maintain the charge balance, on the other hand, suggests that the calcium of the material is substituted with Cs⁺. A more detailed study on the nature and mechanism of Cs sorption onto tobermorite is in progress.

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